Radioactive wastes

[#]Kotelnikov A.R., Tikhomirova V.I., Gurbanov A.A., Chichagov A.V. Ushakovskaya T.V., Kovalsky A.M. Synthesis and study of Mo- and W- bearing sodalites

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key words: [sodalites, solid solutions, experimental study]

Introduction. Creation of matrix materials for radionuclides immobilization on the basis of mineral solid solutions is an actual task of experimental mineralogy. For the first time the opportunity of application of sodalites as matrixes for fixing elements of radioactive waste was shown by Pentinghaus et al. [1]. Then the synthesis iodinebearing sodalite ceramics was carried out in work [2]. The study of leaching resistance has shown high stability of sodalite solid solutions and opportunity of their use as matrix materials. The purpose of the present work was to study molybdenum and tungsten incorporation in sodalites and the possibility of creation of matrixes on the basis of Mo- and W- bearing sodalites.

Synthesis of Mo- and W- bearing sodalites

Starting materials. For the synthesis of Mo- and Wcontaining sodalites and their solid solutions gels of nepheline (NaAlSiO₄) were used. Natural chlorine containing sodalite (2-4 mg on 100-150 mg gel mixture) were added to starting material. For solution preparation in experiments the distilled water and following salts of high purity (Na₂MoO₄*2H₂O and Na₂WO₄*2H₂O) were used.

Run procedure and analytical techniques. The experiences were carried out in the welded platinum tubes (5mm in diameter and 600 mm³ volume). Starting materials (gel mixtures, salts and distilled water) were loaded into the tubes. The concentration of starting solution was 40-60wt.%. Tubes were put into cold seal hydrothermal vessels with external heating. The impermeability of the tubes before and after runs was controlled by weighing.

The experiments were carried out at 450-800°C and $P_{\rm H2O}$ =1-2 kbar. The accuracy of temperature controle was

[#] This study was supported by the RFBR grant (project 00-05-64680).

not worse than 5° C and pressure - 50 bars. The duration of experiences was 10-35 days.

The unit cell parameters were refined for all synthesized sodalites. Samples were recorded in the continuous scan mode of the full X-ray diffraction profile on a PC/HZG-4 automated diffractometer. Angular positions of reflections were estimated by the Spectr-8 program (A.V.Okhulkov, IEM RAS) with correction by the internal standard (spectrally pure Si, a=5.4305 A). Unit cell parameters were refined by 12-17 reflections in the angular range of 7-39 degrees (θ) by the LCC, PUDI [3, 4], REFLAT [5] programs. The chemical compositions of solid products were analysed by the "Camebax" microprobe with energydispersive detector Link AN-10000. The ZAF- correction procedure was used for the calculation of the solid solutions compositions. The accuracy of microprobe analysis was not worse than 2.5 mol.%. The DTA study was carried out for a number of synthesized samples.

Results of experiments. The molybdenum and tungsten- containing sodalites and their solid solutions were synthesized in the hydrothermal conditions at 450-800°C and P_{H2O} =1-2 kbar. The run conditions and the results of X-ray study are given in Table 1.

The DTA study shows that there is a weight loss (in sodalites synthesized at 450°C) during heating which terminates at 185°C. This effect is connected with the loss of water (1.6 wt.%) evidently. The structural transformation connected to allocation of heat exists above 720°C. As it is seen from data of Table 1 the unit cell parameters of Mo- containing sodalites synthesized at low temperature are a little bit higher than for ones synthesized at high temperatures. This effect is likely connected with incorporation of water in sodalite structure. There is no weight loss during heating the sodalites synthesized at 800°C as it is shown by DTA study. It confirms the absence of water in these sodalites.

The solid solutions of (Mo,W)- bearing sodalites are synthesized under hydrothermal conditions at 800°C and $P_{\rm H2O}$ =2 kbar. The synthesized samples are characterized by high crystallinity and their compositions are close to stoichiometric formula. The crystallochemical formula of the synthesized solid solutions of the (Mo,W)- bearing sodalites and the refined unit cell parameters are presented in Table 2.

Table 1. Conditions of synthesis and unit cells parameters of Mo and W- containing sodalites.

Run no.	Phase	T, ℃	P, kbar	Run duration, days	a, [A]	$V, [A]^3$
4899	Mo- Sod	450	1	35	$9.142(1)^{1}$	764.0(1)
5001	Mo- Sod	450	1	16	9.142(1)	764.1(1)
4921	Mo- Sod	700	2	11	9.140(1)	763.4(1)
5022	Mo- Sod	800	2	10	9.137(1)	762.8(1)
5121	Mo- Sod	800	2	10	9.137(1)	762.9(1)
5125	W- Sod	800	2	10	9.148(1)	765.7(1)
-	Mo- Sod ²⁾	-	-	-	9.125	759.5
-	W- Sod ²⁾	-	-	_	9.132	761.6

1) standard errors are given in parentheses and refer to the last decimal place; 2) the PDF-2 data. **Table 2.** The crystallochemical formula and unit cell parameters of the of the (Mo,W)- bearing sodalite solid solutions synthesized from gel mixtures at 800°C and $P_{H2O}=2$ kbar.

Run no.	Crystallochemical formula (calculation to 28 oxygen atoms)	X _W ^{Sod 1)}	a, [A]	$V, [A]^3$
5121	$Na_{7.71}[Al_{5.98}Si_{6.14}]_{12.12}(MoO_4)_{0.97}O_{24.12}$	0	$9.137(1)^{2}$	762.9(1)
5122	$Na_{8.02}[Al_{5.96}Si_{6.16}]_{12.12}(Mo_{0.73}W_{0.18}O_{3.64})O_{24.36}$	0.21	9.140(1)	763.7(1)
5123	$Na_{8.03}[Al_{5.89}Si_{6.14}]_{12.03}(Mo_{0.53}W_{0.39}O_{3.84})O_{24.16}$	0.40	9.142(1)	764.1(1)
5124	$Na_{7.95}[Al_{6.00}Si_{6.11}]_{12.11}(Mo_{0.24}W_{0.62}O_{3.72})O_{24.28}$	0.72	9.143(1)	764.3(1)
5125	$Na_{8.24}[Al_{5.96}Si_{6.10}]_{12.06}(WO_4)_{0.97}O_{24.36}$	1.00	9.148(1)	765.7(1)

1) mole fraction of W in sodalite solid solution; 2) standard errors are given in parentheses and refer to the last decimal place.

Concentration dependences of unit cell parameters of the (Mo,W)- bearing sodalite solid solutions are approximated by the following equations: $a = 9.1369 + 0.024136*X - 0.0428783*X^2 +$ 1 V.I. Vernadskii's Institute of Geochemistry and Analytical Chemistry, Russian Academy of Science

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 $0.0297945^{*}X^{3}$ [A] (±0.001),(1) V = 762.9 + 6.5134^{*}X - 12.7779^{*}X^{2} + 9.074614^{*}X^{3} [A]³ (±0.1),(2)

where X - mole fraction of W in (Mo,W)- sodalite solid solution.

The parameters of Margules model (W1 and W2) for the description of excess volume of (Mo,W)- bearing sodalite solid solutions are -4.38(17) cm³/mol and 2.19(9) cm³/mol accordingly. Excess volume is described by the equation:

 $V^e = X^*(1-X)^*[-4.38^*X + 2.19^*(1-X)]$ (cm³/mol),(3),where X - mole fraction of W in (Mo,W)-sodalite solid solution.

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Stepanets O.V.¹, Komarevskii V.M.¹, Borisov A.P.¹, Spiridonov M.A.² Some aspects of ecological and geochemical studies of the Baltic Sea.

Among ecological problems of the Baltic Sea, two are the most serious. The first problem is related to significant pollution of the sea caused by the Chernobyl nuclear power plant accident in 1986, the second is connected with possible depressurization of chemical weapon sunk in 1946 - 47 years and input of poison compounds into environment.

The report presents results of ecological studies in the Baltic Sea performed in 1996 and 1997 years on the scientific ship "Academician Boris Petrov" of the Institute of Geochemistry and Analytical Chemistry, Russian Academy of Science.



Parameters of technogenic pollution in a surface layer of sea water within all studied sea area, as well as in water and bottom sediments in local areas, in places of chemical weapon storage, are exemplified by 40 stations located in the Baltic Sea and south-eastern part of the North Sea.

Data of radiochemical studies indicate, that horizontal distribution of Cs-137 is characterized by spatial fluctuations, whose presence is assigned firstly by irregular input of radionucleides in 1986 year into different parts of the Baltic Sea and absence of intensive water redistribution between local sea areas (Fig. 1).

Vertical distribution of Cs-137 in depth of a drill core is characterized by a decrease of its activity with depth. However, appreciable variations of activity on different depths were found. That can be related to different rate of sediment movement in these regions. Comparison of the obtained data on the specific activity of Cs-137 in local areas (Table 1) with previous results shows, that by now the horizontal movement of water mass caused dissipation of "patches" of high activity of radionucleides in the subsurface layer of sea water, observed in 1986 and 1987 years. It, in turn, caused throughout decrease of the content of this radionucleide in water. Owing to notable encloseness of the sea, the present concentration of Cs-137 in the Baltic Sea is appreciably higher than that of the adjacent North Sea, although radioactivity of Cs-137 in the North Sea was higher than the level of nuclear pollution of the Baltic Sea before the Chernobyl accident. In contrast to the distribution of Cs-137, concentration of Sr-90 did not significantly change and structure of its concentration fields is similar to that before 1986 year.

Table 1

I ubic I.				
Years	1986	1987	1989	1990
Baltic Sea regions:				
Central part South-eastern	134 <u>±</u> 40	131 ± 40	115 ± 45	125 ± 49
part	162 ± 48	144 ± 66	145 ± 57	130 ± 52
Water area a within Russia	107 - 70	170 - 50	126 + 40	116 - 45
within itussia	186 ± 60	$1/0 \pm 50$	126 ± 49	116 ± 45

Insignificant rate of self-purification of the sea area can be explained both by shallowness of the Baltic Sea and by some geochemical features of bottom sediments. In the Baltic Sea, which is the sea of a positive water balance, water supply into basin, located in humid zone, by sediments and river drainage is higher than the loss by vaporization. Development of halocline at some conditions accounting for vertical and horizontal gradients of salinity, can result in anoxia. In turn, it results in changing in redox conditions and appearance of hydrogen sulfide. As a result, decrease of sorption ability of sediments with respect to nuclear isotopes and equilibration of exchange of radionucleids between sediments and solid phase can occur.

Distribution of chemical parameters in the sub-surface layer of water is regular enough, with local variations, which firstly correspond to geographic location and hydrophysical conditions in the regions of probing. Significant variations of different parameters are not identified in sub-surface waters in the regions of storage of poison compounds in comparison to the "phone" areas.

Table 2.	•
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Excess concentration of chloride ions (Table 2) were found in probes of bottom water in the regions of storage of poison compounds in the Gotland and the Borhholm areas. It implies the presence of poison compounds (phosgene, in our case) in this region and, as a result, products of its hydrolysis:

 $COCl_2 + 2H_2O = H_2CO_3 + 2HCl.$

The deeper horizons of the drill core are the most informative for bottom sediments (Table 3). According to x-ray fluoresence spectral analysis, the significant excess of arsenic, which is a constituent of lewsite, is identified for the Liepäja storage on the level 10-11. The excess content of elements, which are constituents of weapon bodies, is characteristic for the Bornholm storage.

Table also shows, that sub-surface layers of sediments do not show significant excess of technogenic pollution. Such excess does not also appear in bottom and mud waters (the results are obtained on the ship using the Reznikov's hydrochemical laboratory).

The obtained data show, that the revealed pollution of bottom sediments are local enough and are not identified in the upper layers of sediments and in bottom water, at least, in the studied regions. An analysis of the obtained results, accounting for geochemical data on the rate of sedimentation in the Baltic Sea, allows to assume, that chemical weapon, stored in the Liepäja and the Bornholm regions after the war by sinking of separate shells onto bottom, can not promote a loading on an ecosystem in the present time. It is attributed to the screen influence of covering surface of bottom sediments, as well as to sorption of different elements, firstly, hydrolysized in sea water, and nuclear isotopes by muddy surface of bottom sediments. Moreover, the products of hydrolysis of poison compounds, which are simple inorganic compounds, are not toxic.

Another situation can be observed in places of storage of poison compounds in the Skagerrak and the Kattegat straits. In the case of possible depressurization of shells, the spontaneous input of poison compounds can take place. That can lead to significant pollution of environment. However, final conclusions about ecological situation in these regions can be done only after natural investigations.

№	№	Station location	Horizon	pН	Cl,	F,
	St.				g/l	g/l
1	33	Gotland depression,	Surface layers, Bottom	8.2	5.39	3.48
		Liepäja storage	water	7.12	8.29	3.27
2	34	Gotland depression.	Surface layers. Bottom	7.4	5.41	3.35
		"phone" station	water	7.51	6.77	3.48
3	36	Borhholm storage	Surface layers. Bottom	7.91	6.10	3.5
			water	7.33	11.26	4.12
4	37	Borhholm storage	Surface layers. Bottom	7.97	6.81	4.9
			water	7.29	11.69	2.05
5	38	Borhholm storage	Surface layers. Bottom	7.41	6.56	3.08
			water	7.24	6.63	3.15

Table 3.																	
Station location	Station	Layer	Sr	Rb	Pb	As	Br	Zn	Cu	Ni	Co	Cr	V	Ba	Fe	Mn	Ti
	number	(cm)															
	(depth, m)																
Gotland depression,																	
"phone" station	34	10-11	100	160	78	26	320	170	<10	48	26	100	42	0.06	3.4	0.03	<0.0
-														9		0	1
Gotland depression.																	
Liepäja storage	33	10-11	83	220	86	100	230	260	<10	35	36	98	36	<0.0	4.5	0.05	0.35
	(120)													1		6	l
Borhholm storage																	
_	36	10-11	210	190	86	20	270	250	<10	67	33	94	95	0.07	4.7	0.31	<0.0
	(96)																1
Borhholm storage														0			
_	37	10-11	130	98	64	<10	230	240	<10	36	<5	25	62	.046	4.7	0.29	0.3
	(90)																l
Borhholm storage														0	0		
	38	10-11	280	65	65	<10	22	53	<10	18	<5	29	63	.057	.95	0.01	0.16
	(sand)																