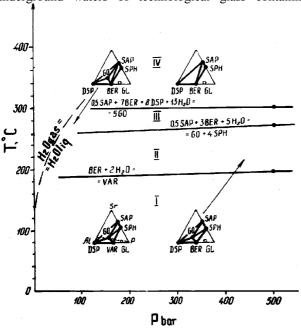
# **Radioactive wastes**

Kotova N.P., Ivanov I.P. Phase equilibria in the system (Na,  $Cs)_2O$ -SrO-Al $_2O_3$ -P $_2O_5$ -H $_2O$  within the range of T=25-350 $^{\circ}$ C and P $_{sat}$  - 500 bar (in connection with the problem of deep radioactive waste repository).

key words [Sr phosphates glass equilibrium solutions]

The system (Na, Cs)<sub>2</sub>O-SrO-Al<sub>2</sub>O<sub>3</sub>-P<sub>2</sub>O<sub>5</sub>-H<sub>2</sub>O has been designed to model phase equilibria at leaching by underground waters of technological glass containing



The T-P phase diagram of this system has been calculated for the range of T=25-350°C and P<sub>sat</sub>=500 bar (fig.1). As follows from the figure equilibria of hydration reactions 1,9, and 4 realized in the T and P range in question are temperature-governed. Pressure is of minute effect for them. Reactions 1,4, and 9 partition the diagram into 4 divariant fields - I, II, II, and IV. Ternary phase associations within each field are illustrated on the composition - paragenesis diagrams. The hydration reactions have shown that in the range of T=25-350°C and P=P<sub>sat</sub>-500 bar the completely stable phases are SAP, SPH, DSP. Variscite (VAR) is stable in the confines of field I at T<186°C. Above of 186°C VAR gets dehydrated (by reaction 1) and transforms to berlinite (BER). At the II=III fields boundary the association GO+SPH transforms to the association SAP+BER by reaction 2, at the III-IV fields boundary GO decomposes completely (following reaction 4) with formation of the association DSP+BER+SAP. At T>300°C 2 ternary associations are DSP+BER+SAP and BER+SAP+SPH.

In the closed system SrO-Al<sub>2</sub>O<sub>3</sub>-P<sub>2</sub>O<sub>5</sub>-H<sub>2</sub>O we have calculated the combined solubility of ternary parageneses of solid phases including SAP, GO, and SPH. With the

radioactive isotopes <sup>90</sup>Sr and <sup>137</sup>Cs under deep repository conditions.

It has been found that phosphates of Na and Cs are readily leached from glass. In the repository site there remain low-soluble phases: Sr-hydroxidapatite (SAP) -  $Sr_{10}(PO_4)_6(OH)_2$ , goyuazite (SO) -  $SrAl_3(PO_4)_2(OH)_5 \cdot H_2O$ , strontium hydrophosphate (SPH) -  $SrHPO_4$ , diaspore (DSP) - AlOOH, berlinite (BER) -  $AlPO_4$ , variscite (VAR) -  $AlPO_4 \cdot 2H_2O$  that enter the system  $SrO-Al_2O_3-P_2O_5-H_2O$ . A thermodynamic database has been created for these phases concluding brine particles.

**Fig.1. T-P phase diagram of the closed system** 1-3 monovariant lines, I-IV - divariant fields, GL-figurative technological glass composition

phase ratios 1:1:1 the compositions of eutonic solutions have been calculated at 25, 186, 260 and 300°C and P<sub>sat</sub> corresponding to phase transition boundaries VAR = BER, GO+SPH=BER+Sap and GO=BER+SAP+DSP on the T-P diagram of the system. It has been found that out of Srphases the minimal solubility is demonstrated by SAP (n·10<sup>-7</sup> mole/kg H<sub>2</sub>O), the maximal one is demonstrated by SPH ( $n\cdot 10^{-4}$  mole/kg H<sub>2</sub>O), so is in the intermediate position between them. The results of calculation are listed in the table. On all the composition - parageneses diagrams the figurative technological glass (GL) composition is plotted. It appears to be shifted far from the region of compositions of low-soluble phases towards the readily soluble component P<sub>2</sub>O<sub>5</sub>. In the temperature range 25-186°C on the line of saturated water vapor the GL composition lies in the field of the triangle VAR-SPH-P<sub>2</sub>O<sub>5</sub>, above 186°C in the field of the triangle BER+SPH+P<sub>2</sub>O<sub>5</sub>. In either case no crystalline phase of P<sub>2</sub>O<sub>5</sub> exists as it is dissolved in water once it is abundant. The concentration of Sr in the solution at the glass-water interaction is: at 186°C and P<sub>sat</sub> - 1.3·10<sup>-3</sup> mole/kg H<sub>2</sub>O, at  $T=260^{\circ}$ C and  $P_{sat}$  -  $1.3\cdot10^{-4}$  mole/kg  $H_2$ O, at  $T=300^{\circ}$ C and  $P_{sat}$  - 2.4·10<sup>-5</sup> mole/kg H<sub>2</sub>O.

Table. Combined (eutonic) solubility of ternary association of solid phases in the system  $SrO-Al_2O_3-P_2O_5-H_2O$  at  $P=P_{sat}$ 

Association	T <sup>o</sup> , C	Equilibrium solution mole/kg/H <sub>2</sub> O			рН	Precipitation
		Al <sub>tot</sub>	Sr tot	P <sub>tot</sub>		sequence
DSP+SAP+B ER	350	4.94.10 <sup>-6</sup>	4.54.10 <sup>-7</sup>	1.99.10 <sup>-5</sup>	5.093	SAP→DSP→BER
DSP+SAP+B ER	300	3.56.10 <sup>-6</sup>	6.56.10 <sup>-6</sup>	4.04.10 <sup>-5</sup>	4.746	SAP→DSP→BER
BER+SAP+SP H	350	9.28.10 <sup>-8</sup>	2.20.10 <sup>-6</sup>	6.10.10 <sup>-4</sup>	4.264	SAP→BER→SPH
BER+SAP+SP H	300	1.74.10 <sup>-7</sup>	2.39.10 <sup>-5</sup>	3.10.10 <sup>-4</sup>	4.101	SAP→.BER→SPH
BER+SAP+SP H	259	2.21.10 <sup>-7</sup>	6.58.10 <sup>-5</sup>	2.69.10 <sup>-4</sup>	4.020	SAP→BER→SPH
GO+SAP+BE R	299	3.34.10 <sup>-6</sup>	6.97.10 <sup>-6</sup>	4.24.10 <sup>-5</sup>	4.732	SAP→GO→BER
GO+SAP+BE R	259	2.27.10 <sup>-7</sup>	6.51.10 <sup>-5</sup>	6.66.10 <sup>-4</sup>	4.208	SAP→GO→BER
DSP+SAP+G O	299	3.53.10 <sup>-6</sup>	6.67.10 <sup>-6</sup>	3.97.10 <sup>-5</sup>	4.755	SAP→DSP→GO
DSP+SAP+G O	259	2.62.10 <sup>-6</sup>	9.30.10 <sup>-6</sup>	2.43.10 <sup>-5</sup>	5.173	SAP→DSP→GO
DSP+SAP+G O	187	2.37.10 <sup>-6</sup>	8.95.10 <sup>-6</sup>	1.07.10 <sup>-5</sup>	6.107	DSP→SAP→GO
DSP+SAP+G O	25	9.50.10 <sup>-8</sup>	1.12.10 <sup>-5</sup>	8.64.10 <sup>-6</sup>	8.713	DSP→SAP→GO
DSP+GO+BE R	299	3.51.10 <sup>-6</sup>	6.49.10 <sup>-6</sup>	4.07.10 <sup>-5</sup>	4.734	DSP→GO→BER
DSP+GO+BE R	259	1.76.10 <sup>-6</sup>	2.82.10 <sup>-6</sup>	5.21.10 <sup>-5</sup>	4.434	DSP→GO→BER
DSP+GO+BE R	187	3.70.10 <sup>-7</sup>	1.87.10 <sup>-7</sup>	8.15.10 <sup>-5</sup>	4.141	DSP→GO→BER
BER+GO+SP H	258	2.13.10 <sup>-7</sup>	6.77.10 <sup>-5</sup>	2.75.10 <sup>-4</sup>	4.196	BER→GO→SPH
BER+GO+SP H	187	5.06.10 <sup>-9</sup>	5.75.10 <sup>-4</sup>	2.30.10 <sup>-3</sup>	3.445	BER→GO→SPH
DSP+GO+VA R	186	3.62.10 <sup>-7</sup>	1.83.10 <sup>-7</sup>	8.09.10 <sup>-5</sup>	4.143	DSP→GO→VAR
DSP+GO+VA R	25	2.95.10 <sup>-9</sup>	2.09.10 <sup>-7</sup>	1.91.10 <sup>-6</sup> '	5.807	DSP→GO→VAR
VAR+GO+SP H	186	4.86.10 <sup>-9</sup>	5.85.10 <sup>-4</sup>	2.33.10 <sup>-3</sup>	3.444	GO→SPH→VAR
VAR+GO+SP H	25	6.76.10 <sup>-12</sup>	3.77.10 <sup>-4</sup>	7.36.10 <sup>-4</sup>	5.601	GO→SPH→VAR
GO+SAP+SP H	258	2.16.10 <sup>-7</sup>	6.70.10 <sup>-5</sup>	2.69.10 <sup>-4</sup>	4.206	SAP→GO→SPH
GO+SAP+SP H	187	1.73.10 <sup>-8</sup>	1.45.10 <sup>-4</sup>	3.23.10 <sup>-4</sup>	4.649	SAP-→GO-→SPH

GO+SAP+SP	25	2.47.10 <sup>-11</sup>	1.16.10 <sup>-4</sup>	$1.74.10^{-4}$	6.870	GO→SAP→SPH
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# Dunaeva A.N. Sorption of <sup>90</sup>Sr on kaolinite, montmorillonite, and illite.

key words [sorption Sr kaolinite illite]

Clays and clay minerals are the important sorbents of heavy metals and radionucleides.

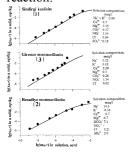
One of the major process of sorptive fixation of radionucleides by clay minerals is their ion exchange with macrocations of clay exchange complex.

It is found, that clay minerals retain their ability for ion exchange within the wide range of pH, from 4 to 10. This interval includes the overwhelming majority of pH of natural waters.

# The purposes of the study:

- 1. To obtain the constants of binary ion exchange of <sup>90</sup>Sr and macrocations K<sup>+</sup>, Na<sup>+</sup>, Ca<sup>2+</sup>, Mg<sup>2+</sup> on kaolinite, montmorillonite, and illite on the basis of treatment of published experimental isotherms of sorption of radionucleides on pure minerals.
- 2. To estimate the dependence of ion exchange sorption on pH.
- 3. On the basis of the obtained data to show the possibility of modeling Sr sorption from multicomponent aqueous solution by polymineral rocks.

<u>Theoretical basis</u> The process of Sr sorption by clay mineral can be described by the following chemical reaction:



$$2A_{1/n} + Sr^{2+} \iff 2Sr_{0.5}X + 2/nA^{n+}$$

where A - exchange cation of mineral-sorbent, n - cation charge, X - clay end-member.

It is suggested in the calculation of the ion exchange equilibria, that only the portion of mineral, which corresponds to its sorptive capacity, is able for reaction. In this case, the constant of reaction accounting for exhausting sorptive capacity, can be written

$$K_{l} = \frac{x_{Sr^{2+}(s,ph.)} \cdot m_{A^{n+}(solut)}^{2/n}}{(C - x_{Sr^{2+}(s,ph.)})^{2/n} \cdot m_{Sr^{2+}(solut)}} \cdot \frac{\gamma_{Sr^{2+}(s,ph.)}^{2/n} \cdot \gamma_{A^{n}(solut)}^{2/n}}{\gamma_{A^{n+}(s,ph.)}^{2/n} \cdot \gamma_{Sr^{2+}(solut)}^{2/n}}$$

where x, m,  $\gamma$  are mole fractions, molalities, and activity coefficients of components. C [eq/kg] - sorptive capacity of minerals, which characterizes the number of positions in sorbents, able for ion exchange. It is suggested, that the sorptive capacity, at which ion exchange occurs, is an ideal solution with respect to sorbed strontium ( $\gamma_{\text{Sr2+(solid)}} = \gamma_{\text{An+(solid)}} = 1$ ). The molalities of ions in an aqueous solution are calculated accounting for hydrolysis process and complex formation in an aqueous phase. Activity coefficient were calculated using the Debye-Hückel theory.

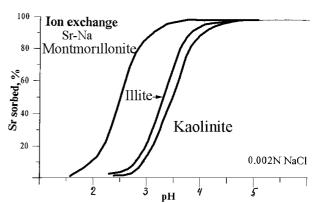


Fig.1. Strontium sorption by clays minerals as a function of pH.

Fig.2. Experimental data of strontium sorption by nature clays [2,3] and sorption isoterms, calculated using K1, K2 values, obtained from Wahlberg J.S. experimental work [1] • - exp. data; — - calculation curves

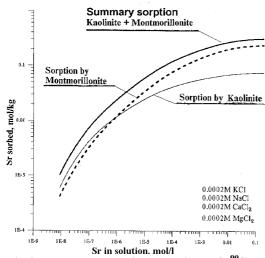


Fig.3. Thermodynamic modeling of <sup>90</sup>Sr sorption from multicomponent solution by polymineral soil fraction

Along with the binary ion exchange sorption of  $Sr^{2+}$ , the competing sorption of proton by clay minerals becomes significant in acid solutions:  $nA_{/n}X+nH \Leftrightarrow nHX+A^{n+}$ .

$$K_2 = \frac{(C - x_{A^{n^+}(s,ph.)} - x_{Sr^{2^+}(s,ph.)})^n \cdot m_{A^{n^+}} \cdot \gamma_{A^{n^+}}}{x_{A^{n^+}(s,ph.)} \cdot m_{H^+}^{n^+} \cdot \gamma_{H^+}^{n^+}}$$

Combination of the equations for  $K_1$  and  $K_2$  gives the equation for dependence of concentration of sorbed strontium on its content in a solution for each *i*-point on the experimental isotherms. In case of the heterovalent sorption such equation turns to:

$$x_{Sr_{s,ph}^{2+}} = \frac{C \cdot K_1}{(K_2 \cdot m_{H^+} + m_{A^+})^2 \cdot \gamma}, \text{ where } \gamma = \frac{\gamma_{A_{solut}^+}^2}{\gamma_{Sr_{solut}^{2+}}}$$

Equation (1) approximates the experimental isotherm of sorption. The function

$$F = \sum_{i=1}^{n} (f_i - x_{Sr_{s,ph.}^{2+}})^2$$

where N - number of experiments,  $f_i$  - concentration of the sorbed Sr in a experiment, is minimized with respect to two unknown parameters,  $K_1$  and  $K_2$ . Sorptive capacity is taken equal to its experimentally obtained value [1].

The experimental isotherms of <sup>90</sup>Sr sorption from K, Na, Ca, Mg-chloride solutions by kaolinite, montmorillonite, and illite, saturated with these components, were treated by the described method. Electrolyte concentration varied from 0.001 to 0.2 N, the interval of <sup>90</sup>Sr was 10<sup>-10</sup> - 0.1 N. Size of clay particles did not exceed 1 mkm. Water: rock ratio was equal to 100: 1.

In case of the heterovalent sorption the experiments were performed in acid solutions (pH = 3). It allowed to estimate the constants for ion exchange H-Na, H-K, and, subsequently, to calculate the constant for H-Ca and H-Mg for above minerals according to the Gess law

Calculations of sorption equilibria by the program SlgSol (Mironenko, 1992) performed on the basis of the

obtained thermodynamic information on constants of binary ion exchange sorption, i.e.  $K_1$  and  $K_2$ , allowed:

- 1) to numerically reproduce the initial experimental data [1] with accuracy up to 10 %;
- 2) to show, that at pH > 4 Sr sorption does not depend on pH of a solution (Fig. 1); therefore, absence of correction for the competing proton sorption does not appreciably influence thermodynamic modeling of <sup>90</sup>Sr behavior in surface and ground waters;
- 3) by constants  $K_1$  and  $K_2$ , calculated from experimental data [1], to reproduce the  $^{90}$ Sr sorption in other model systems, including natural minerals and multicomponent aqueous solutions (Fig. 2);
- 4) to show the possibility of modeling radionucleid sorption from multicomponent solution on polymineral soil (Fig. 3).

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